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Tree-ring $\delta^{13}\text{C}$ tracks flux tower ecosystem productivity estimates in a NE temperate forest

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Abstract

We investigated relationships between tree-ring $\delta^{13}\text{C}$ and growth, and flux tower estimates of gross primary productivity (GPP) at Harvard Forest from 1992 to 2010. Seasonal variations of derived photosynthetic isotope discrimination ($\Delta^{13}\text{C}$) and leaf intercellular CO_2 concentration (c_i) showed significant increasing trends for the dominant deciduous and coniferous species. $\Delta^{13}\text{C}$ was positively correlated to growing-season GPP and is primarily controlled by precipitation and soil moisture indicating that site conditions maintained high stomatal conductance under increasing atmospheric CO_2 levels. Increasing $\Delta^{13}\text{C}$ over the 1992–2010 period is attributed to increasing annual and summer water availability identified at Harvard Forest and across the region. Higher $\Delta^{13}\text{C}$ is coincident with an enhancement in growth and ecosystem-level net carbon uptake. This work suggests that tree-ring $\delta^{13}\text{C}$ could serve as a measure of forest GPP and be used to improve the calibration and predictive skill of ecosystem and carbon cycle models.

 Online supplementary data available from stacks.iop.org/ERL/9/074011/mmedia

Keywords: tree rings, photosynthetic discrimination, carbon cycle, productivity, soil moisture

1. Introduction

Our ability to project the future carbon cycle is limited by our lack of understanding of terrestrial carbon (C) cycle dynamics and the feedbacks that constrain C budgets. Projections of C-flux and C-sequestration from current coupled terrestrial carbon cycle models give widely divergent results (Friedlingstein *et al* 2006). The lack of agreement among projections is, in part, related to poorly constrained model parameters.

Constraining carbon budget projections is an important issue given the current mitigation of anthropogenic emissions by terrestrial ecosystems sequestering C (Le Quéré *et al* 2009). Terrestrial ecosystems sequestered approximately 30% of anthropogenic emissions from 2000 to 2006, and there is interest in the potential to manage forest ecosystems to increase the strength of the carbon sink (Hurt *et al* 2002, Canadell *et al* 2007, Woodbury *et al* 2007). Longer periods of observational data on C-cycle variability, in particular, would help constrain model parameters and improve model performance. CO_2 flux measurements between ecosystems and the atmosphere from eddy covariance flux towers (Baldocchi 2003) are valuable for model evaluation (Suzuki and Ichii 2010, Richardson *et al* 2012, Schaefer *et al* 2012, Keenan



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et al 2012a, Raczka *et al* 2013), but they only measure CO₂ flux conditions over short periods (maximum 20 years, most ≤10 years) in a small footprint (1 km²) and the regional tower density is low (e.g., five sites in northeastern US forests with >10 years of data). Consequently, a variety of other types of measurements are needed to constrain C-cycle model parameters and to improve projections of interannual variability in C-cycle dynamics in terrestrial ecosystems (Pan *et al* 2006, Sims *et al* 2006, Randerson *et al* 2009, Babst *et al* 2013).

Annual measurements of forest carbon production can be obtained from tree rings that integrate the influence of climate and site factors on forest growth (Babst *et al* 2013). Attempts to associate tree-ring width with net ecosystem exchange (NEE) and gross primary productivity (GPP) measurements from flux towers have proven inconclusive (Rocha *et al* 2006). Integrals of annual growth over a few to several years have yielded similar net carbon fluxes as continuous flux tower measurements, but annual increment and flux tower measurements do not show strong inter-annual correlations (Barford *et al* 2001).

The composition of carbon isotopes ($\delta^{13}\text{C}$) of an annual tree ring records the proportions of assimilated carbon and is closely linked to photosynthetic capacity and stomatal regulation during tree growth (Farquhar *et al* 1989, Ogee *et al* 2009). Therefore, the distribution of stable carbon isotopes within a tree ring reflects the carbon assimilation and environmental conditions experienced by a tree through the growing season. The $\delta^{13}\text{C}$ is widely used to study the linkage between environmental conditions and the physiological processes that control tree growth at inter- and intra-annual time resolutions (Loader *et al* 1995). Tree rings, particularly wood cellulose $\delta^{13}\text{C}$, contain valuable records of past climate, leaf-gas exchange, and carbon allocation within trees (e.g. Barbour *et al* 2002). Several studies used tree-ring $\delta^{13}\text{C}$ to reconstruct the intrinsic water use efficiency (iWUE) of trees (e.g. Duquesnay *et al* 1998), and $\delta^{13}\text{C}$ within a tree ring has been shown to track the physiological changes at the canopy level as recorded by the flux towers (Walcroft *et al* 1997, Michelot *et al* 2011). This approach has shown the potential to better understand the link between canopy-level physiology, tree-ring isotopic signature and climate drivers (Offermann *et al* 2011).

In this paper, we identify the relationships between interannual variability in tree-ring $\delta^{13}\text{C}$, tree-ring increment, flux tower measurements of CO₂, and climate for a temperate forest in the northeastern United States. The period of analysis spans 18 years (1992–2010) and is from the longest available record of continuous flux tower measurements (Harvard Forest, Petersham, Massachusetts; Urbanski *et al* 2007). We analyzed the $\delta^{13}\text{C}$ (of α -cellulose) from tree-ring latewood (LW) because it is less dependent on carbon stored during the previous growth year and therefore corresponds to the isotopic signal of recently assimilated C during the growing season. We conducted this analysis for one broad-leaf deciduous and one evergreen needle-leaf co-dominant species in the flux-tower footprint to contrast species response. Specifically, we: (1) identified the relationship between tree-ring $\delta^{13}\text{C}$, GPP, and tree growth, and, (2) assessed the responses of tree physiology and growth to rising atmospheric CO₂ mole

Table 1. Sample characteristics for *Quercus rubra* and *Tsuga canadensis* from the Harvard Forest site.

Species	<i>Quercus rubra</i>	<i>Tsuga canadensis</i>
No. of dated cores	42 (16 trees)	55 (22 trees)
No. of isotopic cores	5 (5 trees)	4 (4 trees)
Mean intercorrelation RW	0.67	0.61
Mean intercorrelation $\delta^{13}\text{C}$	0.71	0.67
EPS RW (1992–2010)	0.95	0.96
EPS $\delta^{13}\text{C}$ (1992–2010)	0.85	0.86

Numbers of dated cores and trees for chronology and isotopic analyses. Mean intercorrelation for isotopic cores. Note that only one core per tree was sampled for isotopic analyses as the intra-tree variability is small in comparison with the inter-tree variability (Daux *et al* 2011).

fraction and climate variation. We also discuss how changes in climate conditions and atmospheric CO₂ concentration over the period of record may have modulated growth and forest productivity over the last 18 years.

2. Materials and methods

2.1. Study site and tree-ring chronologies

The forest within the flux-tower footprint (1 km²; Harvard Forest-EMS tower) is mainly deciduous and is dominated by *Quercus rubra* (L.) (northern red oak; 52% of basal area), *Acer rubrum* L. (red maple; 22% of basal area) and *Tsuga canadensis* (L.) (eastern hemlock; 17% of basal area). Increment core sampling focused on canopy dominant trees. To identify interannual variability in tree growth and tree-ring $\delta^{13}\text{C}$, we cored *Q. rubra* and *T. canadensis* with a 5 mm increment borer at 1.37 m stem height (table 1). Two cores were taken from each tree to build ring-width chronologies. Additionally, we cored five and four trees respectively of each species at the same height with a 12 mm increment borer. The large cores were used to analyze the $\delta^{13}\text{C}$ in the LW of the tree rings. A sample of four trees is sufficient to achieve good precision of the $\delta^{13}\text{C}$ sample mean (McCarroll and Loader 2004, Leavitt 2008).

Ring-width chronologies were developed for each species using standard dendrochronological techniques (Speer 2010). For the site, *Q. rubra* mean age was 97 years (range 71–115 years) and *T. canadensis* mean age was 145 years (range 89–221 years). For isotope cores, *Q. rubra* mean age was 103 years (range 87–113 years) and *T. canadensis* mean age was 142 years (range 93–188 years). All ages were estimated from inner ring dates.

Basal area increment (BAI) was calculated for each tree and species using the dplR package in R (Bunn 2008). We used the ‘outside in’ function to convert raw ring-width measurements to BAI based on the diameter of the tree and the width of each ring moving towards the pith of the tree. The method assumes a circular growth pattern. BAI was used instead of ring width as a surrogate of radial growth and carbon gain because it represents more accurately tree annual biomass increment without the need for standardization

(Biondi and Qeadan 2008). A mean BAI value was computed for each species at an annual resolution, averaged over all the trees cored for both carbon isotope analyses and ring-width chronologies (i.e., 16 trees/42 cores for *Q. rubra* and 22 trees/55 cores for *T. canadensis*).

2.2. Stable isotope analyses

We analyzed the LW portion of each tree ring from individual trees over the 1992–2010 period (table 1). The samples were milled using an ultra-centrifugation mill (Qiagen TissueLysorII) and α -cellulose was extracted from each wood sample following the Soxhlet method elaborated by Green (1963) and modified by Leavitt and Danzer (1993). $\delta^{13}\text{C}$ ratios were measured on the CO_2 produced by α -cellulose combustion in a Costech elemental analyzer coupled with a Thermo Delta V-IRMS. The sample accuracy was determined to be $\pm 0.08\%$ (1 σ standard deviation calculated from the average difference between measured and true internal standard, $n = 13$). The isotopic value is expressed in the delta (δ) notation relative to the VPDB (‰ VPDB).

2.3. Calculation of Δ , c_i and *iWUE*

We used $\delta^{13}\text{C}$ to determine carbon isotope discrimination (Δ) by the plant against atmospheric $\delta^{13}\text{C}$, and variation in plant *iWUE*. The Δ describes the isotopic difference between the $\delta^{13}\text{C}$ of air ($\delta^{13}\text{C}_{\text{air}}$) and the plant ($\delta^{13}\text{C}_{\text{plant}}$) and results from the preferential use of ^{12}C over ^{13}C during photosynthesis. Δ is calculated using:

$$\Delta\text{‰} = \left(\delta^{13}\text{C}_{\text{air}} - \delta^{13}\text{C}_{\text{plant}} \right) / \left(1 + \delta^{13}\text{C}_{\text{plant}} / 1000 \right). \quad (1)$$

Records of $\delta^{13}\text{C}_{\text{air}}$ were obtained from Mauna Loa from 1992 to 2002, and from aircraft measurements collected in Worcester, Massachusetts at 500 m above ground and available from 2003 to 2010 (White and Vaughn 2011). The model of Farquhar *et al* (1982) describes isotopic Δ for C_3 plants via:

$$\Delta\text{‰} = a + (b - a)c_i/c_a, \quad (2)$$

where a is the fractionation during CO_2 diffusion through the stomata (4.4‰; O’Leary 1981); b is the fractionation by RuBP carboxylase (27‰; Farquhar and Richards 1984); and c_i and c_a , are the leaf intercellular space and ambient CO_2 concentrations ($\mu\text{mol Mol}^{-1}$), respectively. We used the data of atmospheric CO_2 concentration (c_a) measured at 29 m height on the eddy-covariance tower to calculate c_i from equation (2). The fractionations due to diffusion and carboxylation are constant but additive; therefore, the $\delta^{13}\text{C}$ records variations in c_i as regulated by two main processes: stomatal conductance (c_a/c_i) and photosynthetic assimilation rate (A). The *iWUE* is the ratio of the net photosynthetic assimilation rate (A) and water vapor conductance ($g_{\text{H}_2\text{O}}$) and is described by Ehleringer and Cerling (1995) as:

$$iWUE = A/g_{\text{H}_2\text{O}} = c_a - c_i(1/1.6). \quad (3)$$

The *iWUE* derived from plant isotope data is used to compare photosynthetic properties independent from evaporative demand (Osmond *et al* 1980), and is therefore, often

applied as an indicator of long-term trends in the internal regulation of carbon uptake and water loss of plants (Seibt *et al* 2008). We used summer (June, July and August) values of $\delta^{13}\text{C}_{\text{air}}$ and c_a to calculate Δ , and to reconstruct c_i and *iWUE* for the period corresponding to LW formation.

2.4. Meteorological and CO_2 flux data and data analyses

We used hourly gap-filled meteorological data from the eddy-covariance tower measured by sensors above the canopy at 29 m height (Urbanski *et al* 2007, Harvard Forest Data Archive HF004) to compute monthly mean temperature, vapor pressure deficit (VPD), incident photosynthetically active radiation, relative humidity as well as monthly total precipitation from 1992 to 2010. Hourly gap-filled GPP was used to provide monthly cumulative GPP. We used the Palmer Drought Severity Index (PDSI) for the central Massachusetts climate region from the National Climatic Data Center.

The association between climate variables and tree-ring Δ chronologies were assessed using correlation analyses. The GPP time series during the growing season (from May to October) were compared to tree-ring Δ chronologies for each species. Correlations were calculated between monthly GPP and LW- Δ using the Pearson (pairwise) product-moment correlation. Trends over the 1992–2010 period were estimated using linear regression.

3. Results

We assessed the number of trees required to extract a common reliable climate signal (thus, the degree to which isotopic composition series vary in parallel) using the expressed population signal (EPS, Robertson *et al* 1997, McCarroll and Pawellek 1998). The EPS values calculated using the 1992–2010 isotopic compositions are ≥ 0.85 for two trees from each species. An EPS value of ≥ 0.85 suggests that the sample size is adequate (Wigley *et al* 1984). At the high frequency investigated in this study (year to year), the $\delta^{13}\text{C}$ series display a high common variance within each species (table 1). We assessed the confidence interval (CI) around annual mean values to account for the differences in the absolute isotopic values (McCarroll and Loader 2004). The sample size of 4 *T. canadensis* and 5 *Q. rubra* provided a $\text{CI}_{95\%} = 0.3\%$ and 0.18% respectively. The mean values of 1992–2010 individual series were calculated for all tree cores to produce one main series for each species (Shi *et al* 2011) hereafter called $\delta^{13}\text{C}_{\text{TSCA}}$ for *T. canadensis* and $\delta^{13}\text{C}_{\text{QURU}}$ for *Q. rubra*.

$\delta^{13}\text{C}$ measurements were significantly different between species ($t = 2.02$, $P < 0.05$). The mean $\delta^{13}\text{C}_{\text{TSCA}}$ display higher ratios than $\delta^{13}\text{C}_{\text{QURU}}$, this difference is probably linked to hydraulic conductivity in conifers which is less efficient than in ring-porous deciduous species (Stuiver and Braziunas 1987, McCulloh *et al* 2010) or other factors such as differences in the maximum photosynthetic capacity or leaf morphology (Barbour *et al* 2002). The species specific time series of annual $\delta^{13}\text{C}$ variations are positively correlated with each other ($r = 0.61$ $P < 0.001$). The derived Δ and c_i series for

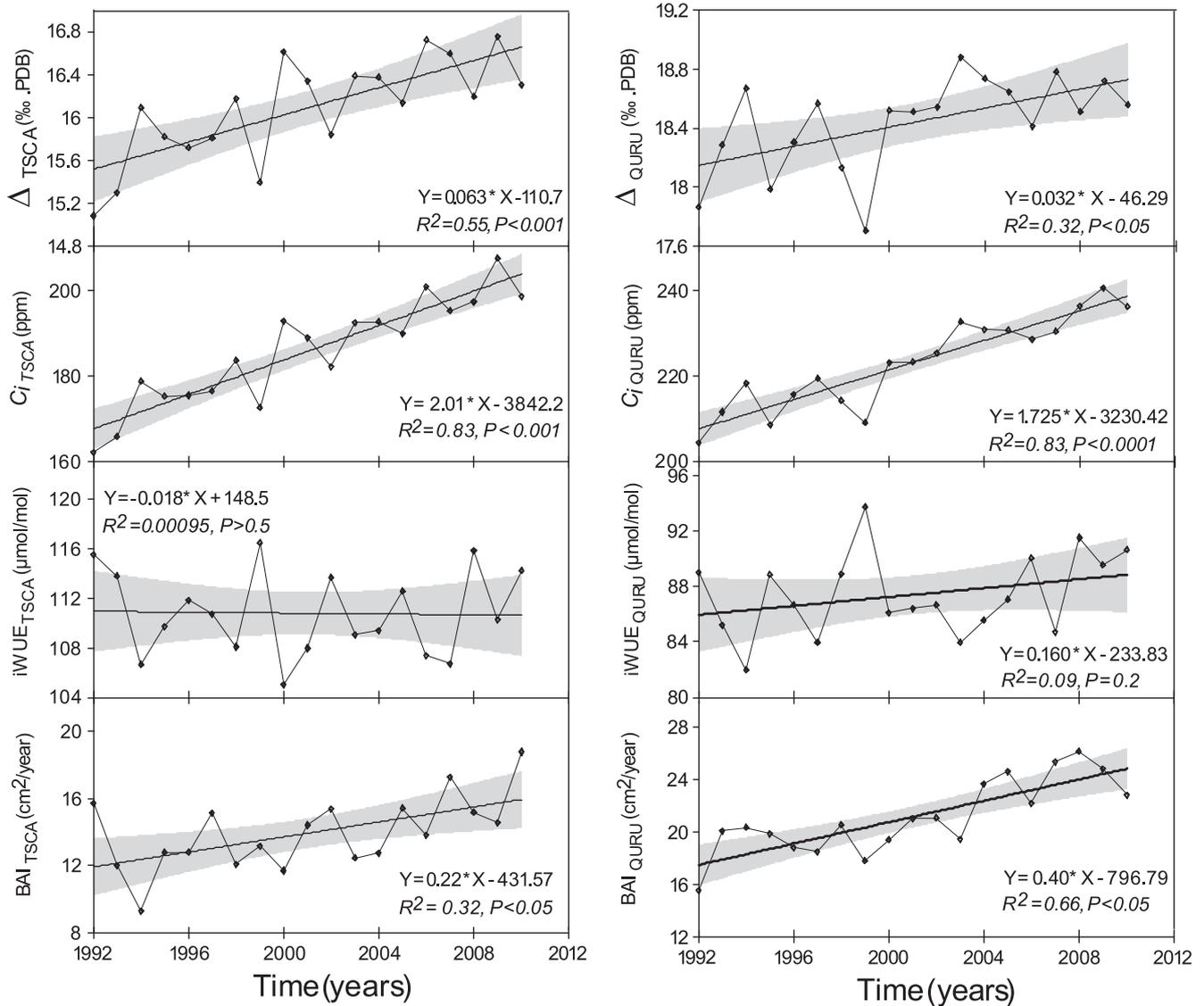


Figure 1. From top to bottom: trends in the $\delta^{13}\text{C}$ discrimination Δ (‰), the leaf intercellular CO_2 concentration (c_i , ppm), intrinsic water use efficiency (iWUE) and basal area increment (BAI) for *T. canadensis* (Left) and *Q. rubra* (Right) throughout time. The slopes of the linear regressions fitted to the data, the coefficients of determination and the P -values are reported for each parameter (see also table 2). The upper and lower levels of the 95% confidence interval for each linear trend are shaded in gray.

each species reveal similar and significant increasing trends over time, although the increase was more pronounced for *T. canadensis* (figure 1). The iWUE over the period of record remained stable which translates into a constant $c_a - c_i$ or an increasing c_i/c_a . A few years did display a higher iWUE in comparison to most years; in particular, iWUE spikes in 1999 in both species records.

The mean BAI also showed a significant increasing trend over the period of record for *Q. rubra* ($R^2 = 0.64$, $P = 0.0001$) but the trend was less pronounced for *T. canadensis* ($R^2 = 0.32$, $P < 0.05$). Overall, the trends analysis using linear regression shows that Δ , c_i and BAI have increased for each species. The changes were significant ($p < 0.05$) and the slopes were positive (table 2).

The Δ was strongly correlated with July (0.55, $P < 0.05$) and October (0.53, $P < 0.05$) GPP for *T. canadensis*, and with July (0.49, $P < 0.05$), August (0.58, $P < 0.01$), and October

(0.55, $P < 0.05$) GPP for *Q. rubra*. Online supplementary table S1 (available at stacks.iop.org/ERL/9/074011/mmedia) summarizes the Pearson correlation obtained between monthly climate variables and Δ_{QURU} and Δ_{TSCA} . Overall, the strongest positive correlations were found with PDSI from May to August for *T. canadensis* and for May for *Q. rubra*. A positive correlation was found between Δ of both species and early-growing season precipitation (April and May).

4. Discussion

4.1. Relationship between Δ , growth, and flux tower estimates of GPP

The carbon contribution to stem growth during LW formation should primarily originate from the immediate product of

Table 2. Trend statistic, per species, for isotopic discrimination (Δ), intercellular CO₂ concentration (c_i), intrinsic water use efficiency (iWUE) and basal area increment (BAI). All units are given per year (yr). The slopes are estimated using linear regression.

	Slope	Sig
Δ_{TSCA} (‰ yr ⁻¹)	0.06	<0.0001*
Δ_{QURU} (‰ yr ⁻¹)	0.03	0.01**
$c_{i,TSCA}$ (ppm yr ⁻¹)	2.01	<0.0001*
$c_{i,QURU}$ (ppm yr ⁻¹)	1.72	<0.0001*
iWUE _{TSCA} (μmol mol yr ⁻¹)	-0.018	0.9***
iWUE _{QURU} (μmol mol yr ⁻¹)	0.16	0.2***
BAI _{TSCA} (cm ² yr ⁻¹)	0.22	0.01**
BAI _{QURU} (cm ² yr ⁻¹)	0.40	<0.001*

The significance of each variable slope is given * $p < 0.01$, ** $p < 0.05$ and *** Positive or negative but non-significant.

carboxylation thus reflecting the $\delta^{13}C$ signature of recent assimilates (Helle and Schleser 2004). The interannual variability of Δ was compared with monthly GPP to identify the seasonality of carbon uptake and allocation to above-ground woody biomass. The correlations shown in figure 2 indicate that recent photosynthates provided the substrates for stem radial growth occurring in the second part of the growing season (Gessler *et al* 2009, Offermann *et al* 2011). The Δ -GPP correlations are logical based on the timing of LW formation for both *T. canadensis* and *Q. rubra*. LW tracheid formation and cell wall thickening in *T. canadensis* begins later in the growing season (late July–October, Skene 1972) coincident with the significant correlations. In contrast, LW formation in *Quercus* species begins approximately five weeks after the unfolding of first leaves and continues until radial growth cessation in late summer or early fall (Zasada and Zahner 1969, Voelker *et al* 2012).

The Δ -GPP correlations are consistent with the seasonality of carbon uptake from eddy flux measurements and C storage estimates at the deciduous forest-EMS tower site and the Hemlock Forest tower site located 0.5 km from the EMS tower (Hadley and Schedlbauer 2002, Hadley *et al* 2009). Net C uptake at the deciduous forest site occurs between the end of May and mid-October. At the hemlock site, the peak in net C uptake and storage occurs in April and May. This period is not influencing the $\delta^{13}C$ in the LW, although it shows the effect of conifers on the annual pattern of the C exchange. Conifers currently represent 17% of the total basal area at the deciduous forest site. After C uptake declines in June, it increases in July, and falls in August before a second uptake peak in October, consistent with the Δ -GPP correlations for *T. canadensis*.

The species seasonality of carbon allocation is further reflected in the relationship between LW- Δ and growth. The correlation for *Q. rubra* LW- Δ and mean BAI was significant ($r = 0.72$, $P < 0.001$), and was slightly higher with the mean *Q. rubra* BAI of all sampled trees ($r = 0.75$, $P < 0.001$). To illustrate the relationship between aboveground carbon gain and photosynthetic discrimination we regressed BAI against Δ , and a strong positive trend was obtained for *Q. rubra*

($R^2 = 0.39$ $P < 0.01$). The growth increase of *Q. rubra* could be interpreted as the result of a greater photosynthetic rate induced by increased c_i concentrations (von Caemmerer and Farquhar 1981, Schubert and Jahren 2012), and a substantial allocation of carbohydrates to LW formation and radial growth (Keel *et al* 2006, Palacio *et al* 2011). However stage of stand development may also be contributing to this increase in BAI (e.g. Foster *et al* 2014, figure S1). A relationship between Δ and BAI was not found for *T. canadensis*, probably because carbon sequestered after June contributes little to the current year radial growth or that carbon allocation differs by species (Richardson *et al* 2013). The trend in the BAI growth enhancement for *Q. rubra* is consistent with changes in the measured annual increment of aboveground biomass in the flux-tower footprint from 1993–2010 (~20% increase, Data Archive: HF069), and provides a quantification of carbon stored in aboveground woody biomass. Interestingly, the aboveground biomass over the last decade accounted for ~50% of the total carbon sequestered (Keenan *et al* 2012a) confirming the relationship we found between *Q. rubra* BAI and Δ , and growing season GPP (figure 2). The remaining carbon uptake could be attributed to litter or soil pools; however, these carbon proportions and stocks are not accounted for in the tree-rings $\delta^{13}C$.

The inter-annual variability and fraction of carbon uptake and sequestration explained by the tree-ring derived Δ (figures 2(c), (d)) provides a quantitative proxy that can be used to constrain models that estimate forest productivity. Process-based models fail to accurately reproduce the observed inter-annual variability of C-fluxes (Keenan *et al* 2012b) mainly because of inaccurate model allocation structure and lagged effects of climate variability on tree growth and physiology (Gough *et al* 2009). In contrast, the Δ integrates biotic factors which modulate the $\delta^{13}C$ fractionation and the isotopic composition of the total pool of carbon fixed during photosynthesis at seasonal and annual cycles (Leavitt 1993).

4.2. Climate drivers and physiological implication of the Δ trend

We examined the climatic drivers of Δ using meteorological data from the eddy-covariance tower and PDSI. The climate signal within both species Δ series is dominated by monthly precipitation and PDSI during the growing season (table S1), although they have a stronger influence on *T. canadensis* due to its particularly drought-sensitive nature caused by a relatively shallow rooting-depth (Cook and Cole 1991).

The carbon isotope discrimination properties can be used to assess how trees are responding to increasing atmospheric concentration c_a and changes in climate. Three physiological response scenarios described by Saurer *et al* (2004) identify possible changes in iWUE and c_i following c_a increase: (1) c_i remains constant such that c_i/c_a decreases and iWUE increases; (2) c_i increases proportional to c_a such that c_i/c_a remains constant and iWUE increases; (3) c_i increases at the same rate as c_a , c_i/c_a increases and iWUE remains constant. In this study, Δ and c_i increased and there was no increase in iWUE for both species (figure 2, table 2). The c_i/c_a calculated for

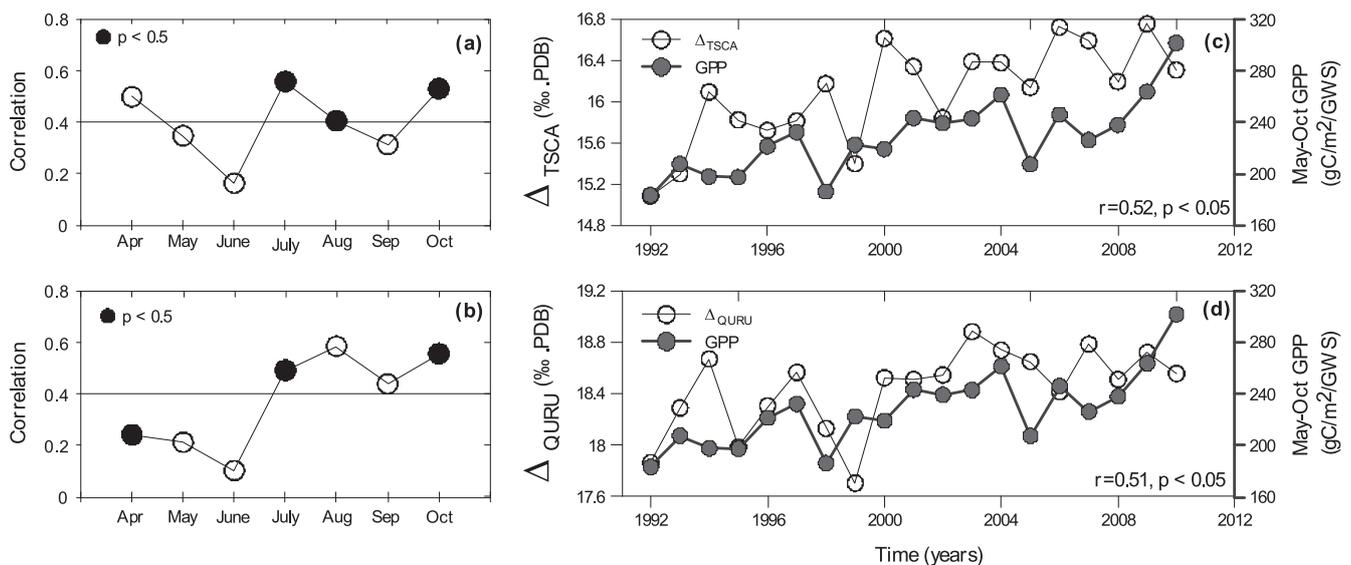


Figure 2. Tree-ring Δ chronologies and monthly GPP data from April to October. Pearson correlations from 1992 to 2010 between Δ and mean monthly GPP data for *T. canadensis* (a) and *Q. rubra* (b). May–October GPP and the Δ chronologies for *T. canadensis* (c) and *Q. rubra* (d) versus time.

both species increases indicating c_i follows the increase in c_a at a rate of $2 \pm 0.21 \text{ ppm yr}^{-1}$ for *T. canadensis*, and $1.7 \pm 0.17 \text{ ppm yr}^{-1}$ for *Q. rubra*. The c_i/c_a increase and no change in $iWUE$ suggest a weak stomatal response to c_a increase and is described as a passive response to changes in c_a (McCarroll *et al* 2009) where neither stomatal conductance nor photosynthetic rate change. The most common response documented in trees is an active response as stomatal conductance is reduced (Bert *et al* 1997, Saurer *et al* 2004, Peñuelas *et al* 2011). The $iWUE$ trends derived from the tree-rings $\delta^{13}\text{C}$ in our study show a divergence from the instantaneous WUE (the ratio of carbon assimilation to transpiration, Farquhar and Richards 1984) trend documented at the ecosystem level using continuous measurements of CO_2 and water vapor fluxes at HF (Keenan *et al* 2013). The canopy-integrated water use efficiency calculated as the ratio between gross ecosystem photosynthesis (GEP) and Ecosystem evapotranspiration (E) shows a significant increase over the measurements period ($3.5\% \text{ change yr}^{-1}$, $P = 0.01$). The latter was calculated taking into account the atmospheric evaporative demand and therefore is referred to as ‘inherent’ water use efficiency (W_{ei}). The ‘intrinsic’ water use efficiency as derived from tree-rings $\delta^{13}\text{C}$ is often used as an indicator of long-term trends in the internal regulation of carbon uptake and water loss independently from evaporative demand (Osmond *et al* 1980). Therefore $iWUE$ inferred from $\delta^{13}\text{C}$ may not be representative of instantaneous or inherent WUE . Although both are affected by similar processes (i.e. stomatal conductance), they can vary independently because they are influenced by additional factors like mesophyll conductance, leaf N-content and C-respiratory losses (Griffiths *et al* 1999, Seibt *et al* 2008). We note that when $iWUE$ is calculated from the flux tower measurements (as the ratio between GEP and canopy water conductance), the magnitude of the trend is lower and the statistical strength is no longer significant

compared to those calculated for W_{ei} ($1\% \text{ change yr}^{-1}$, $P = 0.2$, Keenan *et al* 2013).

The Δ trend found in this study during the last 18 years is unusual but not unique. Results from literature often reported a decrease in Δ inferred from tree-ring records, and thus a strong improvement of $iWUE$ during the last 100 years under increased atmospheric CO_2 (Arneeth *et al* 2002). However, Marshall and Monserud (1996) showed that $iWUE$ has remained static in sites from western US and a switch in $iWUE$ from an active to a passive response around AD 1970 has been noted in several European sites (Gagen *et al* 2011). The spatial variability of stomatal regulation and $iWUE$ response to increasing c_a is explained by the strong dependence of water and carbon usage on moisture stress experienced at a specific site and species sensitivity to moisture availability (Warren *et al* 2001). The increased discrimination is a reflection of the availability of CO_2 . When the level of CO_2 (c_i) is high, large discrimination is observed reflecting the RuBisCO enzymatic preference for $^{12}\text{CO}_2$ (Stewart *et al* 1995). Clearly, the Δ chronologies at HF indicate that despite higher c_a concentrations, factors like physiology and site condition contributed to stomatal openness and conductance.

The strong correlation between Δ and growing season PDSI and precipitation for both species suggest a link between water availability and stomatal conductance (Dupouey *et al* 1993). The climatic data for central Massachusetts indicates that precipitation and moisture (PDSI) are high at HF and are rarely limiting tree growth (Voelker 2011). In fact, the region became wetter over the period of analysis. The mean annual precipitation from 1898 to 2010 was 107.45 cm and increased $0.28 \pm 0.06 \text{ cm yr}^{-1}$ ($p < 0.0001$). The trend in precipitation is also reflected in a trend of increasing PDSI of $0.03 \pm 0.005 \text{ units yr}^{-1}$ ($P < 0.0001$) over the same period. Since 1992, precipitation and PDSI have

increased at rates of $1.30 \pm 0.67 \text{ cm yr}^{-1}$ ($P < 0.07$) and $0.15 \pm 0.05 \text{ units yr}^{-1}$ ($P < 0.006$), respectively. The increase in regional scale water availability (Wang *et al* 2013) is also evident at the HF site. Annual precipitation and growing season precipitation have increased by 10% and 15% over the last 18 years, respectively (Keenan *et al* 2013).

We conclude on the basis of the photosynthetic discrimination sensitivity to water availability that the long-term increase in moisture which controls Δ drove the increase in stomatal conductance and carbon photosynthetic discrimination. A positive correlation between Δ and mean annual and summer precipitation has been confirmed for modern leaf tissues from multiple biomes in North America (Diefendorf *et al* 2010). This correlation is explained by the effect of the mean annual and summer precipitation levels on both soil water status and VPD (Hartman and Danin 2010). It is noteworthy that despite the mesic conditions of the HF site, precipitation and soil moisture levels remain very strong predictors of CO_2 discrimination. Similar to our findings, drought sensitivity has been demonstrated for radial growth of *T. canadensis* and *Q. rubra* growing in mesic sites located in the New York City watershed (Pederson *et al* 2013).

The sensitivity to water availability is further illustrated by iWUE data for the year 1999. Both species exhibited a higher iWUE reflecting lower Δ values (figure 1). The summer of 1999 was very dry with decreased soil moisture levels resulting in the restriction of the CO_2 supply to the leaf (Brugnoli *et al* 1988) because of stomatal closure, and in a weaker carbon uptake observed in the summertime NEE record (Urbanski *et al* 2007).

5. Conclusion

An enhancement in photosynthetic isotopic ^{13}C discrimination and c_i was observed over the last 18 years for two co-dominant species at the Harvard Forest. The combined effect of increased water availability and higher atmospheric CO_2 concentration lead to increased plant CO_2 assimilation. The higher net photosynthesis can in part explain the growth enhancement of *Q. rubra* and above-ground GPP (e.g. Linares and Camarero 2012) and is consistent with the observed higher CO_2 uptake and C storage documented in the flux-tower measurements over the same period. The tree-ring Δ -GPP relationship identified in this study can be used as a quantitative proxy to reconstruct and interpret past forest productivity, as driven by climate variability and in response to long-term atmospheric CO_2 increase.

The photosynthetic carbon isotopic discrimination model is embedded in most of the models of forest carbon cycling (Richardson *et al* 2012) that are coupled with earth-system models to project terrestrial carbon cycle and feedbacks to climate change (Sitch *et al* 2008). We show that the data provided by tree-ring $\delta^{13}\text{C}$, recording the seasonal and interannual patterns of plant carbon assimilation, can be used to inform forest ecosystem model parameterization (Bodin *et al* 2013), to improve simulations of monthly scale ecosystem-function (e.g. Medvigy *et al* 2013), and to constrain

longer term terrestrial carbon dynamics in response to climate fluctuations and forest management (e.g. Brooks and Mitchell 2011).

Whether the documented isotopic trends in the present study are merely temporary or a site specific phenomenon remains to be seen. Further investigations of tree-ring $\delta^{13}\text{C}$ and ecosystem physiology measurements from other flux-tower sites in northeastern US forests need to be explored to assess physiological responses at the regional scale.

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